

Emission of trace gases and aerosols from biomass burning

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Abstract. A large body of information on emissions from the various types of biomass burning has been accumulated over the past decade, to a large extent as a result of International Geosphere-Biosphere Programme/International Global Atmospheric Chemistry research activities. Yet this information has not been readily accessible to the atmospheric chemistry community because it was scattered over a large number of publications and reported in numerous different units and reference systems. We have critically evaluated the presently available data and integrated these into a consistent format. On the basis of this analysis we present a set of emission factors for a large variety of species emitted from biomass fires. Where data were not available, we have proposed estimates based on appropriate extrapolation techniques. We have derived global estimates of pyrogenic emissions for important species emitted by the various types of biomass burning and compared our estimates with results from inverse modeling studies.

1. Introduction

Human evolution and the use of fire have gone hand in hand ever since the origin of our species in the savannas and woodlands of Africa [Schüle, 1990]. As a result, air pollution from the smoke of biomass fires has been humanity's constant companion for some 2 million years, and its ancient impact on human health is reflected in soot deposits in the lungs of mummies. Emissions from fossil fuel burning gained notoriety as air pollutants in medieval England, but only in the last 2 centuries have they begun to play an important role worldwide. Yet the scientific exploration of air pollution has focused initially only on this much more recent threat, and the first pioneering papers on the impact of biomass burning on the chemistry of the atmosphere were only published in the 1970s and early 1980s [e.g., Eagan *et al.*, 1974; Radke *et al.*, 1978; Crutzen *et al.*, 1979]. Scientific interest in this topic grew when early estimates of pyrogenic emissions suggested that for some atmospheric pollutants biomass burning could rival fossil fuel use as a source of atmospheric pollution [Seiler and Crutzen, 1980; Crutzen and Andreae, 1990] and when it became evident that these emissions could affect large areas of the world as a consequence of long-range transport [Andreae, 1983; Kirchhoff and Nobre, 1986; Reichle *et al.*, 1986; Fishman *et al.*, 1990].

Satellite and airborne observations have shown elevated levels of O₃, CO, and other trace gases over vast areas of Africa, South America, the tropical Atlantic, and the Indian and Pacific Oceans [Andreae *et al.*, 1988; Watson *et al.*, 1990; Fishman *et al.*, 1991; Andreae *et al.*, 1994; Cahoon *et al.*, 1994; Folkins *et al.*, 1995; Browell *et al.*, 1996; Connors *et al.*, 1996; Talbot *et al.*, 1996; Thompson *et al.*, 1996; Jonquieres *et al.*, 1998; Novelli *et al.*, 1998; Blake *et al.*, 1999; Gregory *et al.*, 1999; Singh *et al.*, 2000]. Smoke aerosols perturb regional, and probably global, radiation budgets by their light-scattering effects and by their influence on cloud microphysical processes [Crutzen and Andreae, 1990; Penner *et al.*, 1992; Andreae, 1995]. Following well-publicized large fire catastrophes in recent years and intensive scientific efforts over the last decade, the general public as

well as the scientific community are now aware that emissions from biomass burning represent a large perturbation to global atmospheric chemistry, especially in the tropics.

To assess the atmospheric impact of biomass burning, and especially to represent it quantitatively in models of atmospheric transport and chemistry, accurate data on the emission of trace gases and aerosols from biomass fires are required. Emissions must typically be represented in the form of spatiotemporally resolved fields, where the emission per unit area and time is provided at a specified spatial and temporal resolution. These fields are obtained by multiplying an exposure term, for example, the amount of biomass burned within a grid cell during a time interval, with an emission factor, that is, the amount of the chemical species released per mass of biomass burned.

In the last couple of decades the efforts of individual groups to characterize fire emissions have been greatly extended by a number of coordinated biomass burning experiments in various ecosystems throughout the world, often under the auspices of the International Global Atmospheric Chemistry (IGAC) Project of the International Geosphere-Biosphere Programme (IGBP). Examples of these experiments are Dynamique et Chimie Atmosphérique en Forêt Equatoriale-Fire of Savannas (DECAFE-FOS) in West Africa [Lacaux *et al.*, 1995], Southern Africa Fire-Atmosphere Research Initiative (SAFARI 92) [Andreae *et al.*, 1996c; Lindesay *et al.*, 1996], Transport and Atmospheric Chemistry Near the Equator-Atlantic (TRACE-A) over Brazil, southern Africa, and the South Atlantic [Fishman *et al.*, 1996], Experiment for Regional Sources and Sinks of Oxidants in central Africa (EXPRESSO) [Delmas *et al.*, 1999], Smoke, Clouds, and Radiation-Brazil (SCAR-B) in Brazil [Kaufman *et al.*, 1998], and Fire Research Campaign Asia-North (FIRESCAN) in central Siberia [FIRESCAN Science Team, 1996]. These coordinated studies and numerous independent smaller investigations have resulted in a large body of information on the emission characteristics of biomass fires. These data describe qualitatively and quantitatively the pyrogenic emission of chemically and radiatively important gases and aerosol species, but, unfortunately, this information is scattered through the scientific literature and presented in a large variety of formats and units, making its use very difficult. An earlier synthesis of then available data was provided by Delmas *et al.* [1995], but it covers only a minor fraction of the information that has accumulated in recent years.

The objective of this paper is to synthesize the currently available data on fire emission characteristics for a large number of chemical species into a consistent set of units. In contrast to some previous summaries that gave only generic estimates independent of the type of fire [e.g., *Andreae, 1993*], here we provide separate emission factors for the different types of biomass burning, such as deforestation fires in the tropics, savanna fires, etc. We then combine the emission factor data with exposure estimates for the various fire categories to provide global estimates of emissions of biomass burning.

2. Methods

2.1. Definitions

Emission information is represented in two basic forms: emission ratios or emission factors. These parameters relate the emission of a particular species of interest to that of a reference species, such as CO₂ or CO (emission ratio) or to the amount of fuel burned (emission factor); their definition and use will be discussed below.

Emission ratios are obtained by dividing the excess trace species concentrations measured in a fire plume by the excess concentration of a simultaneously measured reference gas, for example, CO₂ or CO. To obtain "excess" concentrations, the ambient background concentrations must be subtracted from the values measured in the smoke. For example, the emission ratio of methyl chloride, CH₃Cl, relative to CO is

$$ER_{\text{CH}_3\text{Cl}/\text{CO}} = \frac{\Delta \text{CH}_3\text{Cl}}{\Delta \text{CO}} = \frac{(\text{CH}_3\text{Cl})_{\text{smoke}} - (\text{CH}_3\text{Cl})_{\text{ambient}}}{(\text{CO})_{\text{smoke}} - (\text{CO})_{\text{ambient}}}$$

Alternatively, the emission ratio can be determined as the regression slope of the species concentration versus the reference species. The various techniques for these calculations and the associated errors are discussed by *Le Canut et al. [1996]*. For gases, emission ratios are usually expressed in terms of molar ratios. For aerosols, emission ratios are most often given in units of mass aerosol per kilogram CO₂-carbon (g kg⁻¹ C(CO₂)). The most commonly used reference species are CO₂ and CO, but other compounds such as CH₄ and acetylene (C₂H₂) have also been used.

Besides practical considerations, for example, the availability of a particular analyzer in a given study, the selection of CO or CO₂ as reference gas is determined by the ultimate objective of the analysis and on the combustion type (flaming or smoldering) from which the species is preferentially released. CO is emitted predominantly by smoldering combustion and is therefore a suitable reference gas for other species emitted mostly during smoldering. Close correlations between smoldering-derived gases and CO can usually be observed, which allows fairly accurate estimation of trace gas emissions from fires for which the CO emission is known. In contrast, correlations of smoldering-derived gases with CO₂ tend to be relatively poor, since the relative proportion of flaming versus smoldering combustion in different fires, or even in different parts of the same fire, results in variable trace gas to CO₂ ratios [*Lobert et al., 1991*]. For smoldering-related species containing nitrogen or halogen elements, for example, NH₃ or the methyl halides, the emission ratio relative to CO is also dependent on fuel composition, i.e., the nitrogen or halogen element content of the fuel.

Flaming-derived species, however, correlate well with CO₂, which makes this the obvious reference gas for these compounds. Because most of the biomass carbon is released from a fire as CO₂, the emission ratio of a gas relative to CO₂ permits the estimation of trace gas emission from fires based on the amount of biomass burned. Therefore this ratio is the most suitable parameter for regional or global estimations, and consequently, even smoldering-

related species are often related to CO₂ as a reference species. This can be done either by calculating an emission ratio directly from the regression slope of the species concentration to that of CO₂ or by first deriving an emission ratio to CO and multiplying that with the average CO/CO₂ emission ratio for the fire.

Emission ratios have the advantage that they only require simultaneous measurements of the species of interest and the reference species in the smoke and appropriate measurements in the background air. They do not require knowledge of the fuel composition and the amounts of fuel burned and are therefore suitable for field studies, particularly airborne plume measurements where fuel information is usually not available.

In principle, it is possible to obtain an emission ratio from a single pair of plume and background measurements. In practice, however, several measurements are usually averaged, or information from continuous analyzers is used in order to derive data representative of a fire or fire phase. Since the emission ratio of smoldering-related species to CO₂ or the ratio of flaming-related species to CO can vary widely through the course of a fire, it is essential that this averaging is done correctly. When the emission ratio for an entire fire is sought, two approaches are possible: (1) The entire amounts of trace and reference species emitted during the fire can be integrated and divided by one another, or (2) an average of instantaneous emission ratios can be calculated, in which the individual measurements are weighted according to the amount of fuel combusted in the time interval represented by the measurement. These approaches require information that is only available in experimental fires in the laboratory or to a limited extent in the field. For aircraft measurements one has to rely on the mixing processes in a plume to perform this averaging. This requires that both flaming and smoldering emissions are entrained into the plume and that the plume is sampled in a representative fashion. Ground-based sampling has a tendency to overrepresent smoldering emissions which tend to be emitted during less vigorous phases of a fire and therefore to remain closer to the ground, while airborne sampling may have a bias toward emissions from the flaming phase that rise to higher altitudes [*Andreae et al., 1996a*]. For the calculation of average emission factors in Table 1 we have assumed that airborne measurements represent an unbiased average of the emissions. When smoldering and flaming emissions were given separately in ground-based studies, we have combined them to represent the complete fire. For this purpose we either used data on the fractions of fuel combusted in the smoldering and flaming stages provided in a given study or, when this information was not available, typical values from other studies on the same type of fire.

Another parameter frequently used to characterize emissions from fires is the emission factor, which is defined as the amount of a compound released (M_x) per amount of dry fuel consumed (M_{biomass}), expressed in units of g kg⁻¹. Calculation of this parameter requires knowledge of the carbon content of the biomass burned and the carbon budget of the fire; both parameters are difficult to establish in the field as opposed to laboratory experiments where they are readily determined. The total carbon released is usually estimated by adding the measured concentrations of CO₂, CO, hydrocarbons, and particulate carbon, when this information is available. The emission factor can then be expressed as

$$EF_x = \frac{M_x}{M_{\text{biomass}}} = \frac{M_x}{M_c} [C]_{\text{biomass}}$$

$$EF_x \cong \frac{[x]}{\sum([C_{\text{CO}_2}] + [C_{\text{CO}}] + [C_{\text{CH}_4}] + [C_{\text{VOC}}] + [C_{\text{aeros}}] + \dots)} [C]_{\text{biomass}},$$

where M_c is the mass of carbon emitted, $[C]_{\text{biomass}}$ is the carbon concentration in the biomass burned, $[x]$ is the concentration of

species x in the smoke, and $[C_{CO_2}]$ etc. are the concentrations of the various carbon species in the smoke. Alternatively, the carbon budget of the fire can be estimated by multiplying the CO_2 in the smoke with an assumed factor to represent non- CO_2 carbon emissions (i.e., CO , hydrocarbons, and aerosol carbon). Where fuel and residue data at the ground are not available, a fuel carbon content of 45% is usually assumed in order to derive emission factors from emission ratios.

2.2. Conversion of Units

We adopted the use of emission factors, as defined in section 2.1, to express all emission data in this paper. The reason for this choice was that emission factors allow the most direct estimation of trace gas and aerosol release when the amount of biomass burned is given. This choice made it necessary to convert data given in other forms into emission factors. We proceeded as follows.

When emission factors were given following the same definition as specified in section 2.1, the data were used unchanged. Where emission factors were given relative to other fuel mass indicators (e.g., as mass species emitted per mass carbon burned or released), we multiplied the data with an appropriate conversion factor, typically, the carbon content of the fuel. When this information was not provided by the authors, we used a default carbon content of 45%.

To convert data given as molar emission ratios, we used the following equation:

$$EF_x = ER_{(X/Y)} \frac{MW_X}{MW_Y} EF_Y,$$

where $ER_{(X/Y)}$ is the emission ratio of species x relative to the reference species Y , MW_X and MW_Y are the molecular weights of the species x and the reference species Y , and EF_Y is the emission factor of the reference species. If EF_Y was provided in the data source, it was used directly; otherwise, the mean EF_Y for the appropriate type of fire (forest, savanna, etc.) was applied.

3. Results and Discussion

3.1. Pyrogenic Emissions and the Combustion Process

The emission factors of the various smoke constituents are determined by the composition of the fuel and by the physical and chemical processes during combustion. Since the carbon content of fuels varies only over a fairly limited range (typically 37–54%), the emission factors for compounds that contain only carbon, oxygen, and hydrogen are predominantly a function of the combustion process. Charcoal burning is an exception to this rule because of the high carbon content of this fuel (>80%). The emission of substances containing minor elements, such as nitrogen, sulfur, and the halogens, is determined both by the concentrations of those elements in the fuel and by the combustion conditions.

The processes during the combustion of biomass have been described in detail [Lobert and Warnatz, 1993; Yokelson *et al.*, 1996, 1997] and will be reviewed here only briefly. Combustion of the individual fuel elements proceeds through a sequence of stages (ignition, flaming plus glowing plus pyrolysis, glowing plus pyrolysis (smoldering), glowing, and extinction), each with different chemical processes that result in different emissions. Plant biomass consists of cellulose and hemicelluloses (typically 50–70% dry matter), lignin (15–35%), proteins, amino acids, and other metabolites, including volatile substances (alcohols, aldehydes, terpenes, etc.). In addition, it contains minerals (up to 10%) and water (up to 60%). Thermal degradation begins with a drying/distillation step, in which water and volatiles are released, followed by pyrolysis, during which thermal cracking of the fuel molecules occurs. This results in the formation of char (less volatile solids of

high C content), tar (molecules of intermediate molecular weight), and volatile compounds in the form of a flammable white smoke. When temperatures in the fuel bed exceed 450 K, the process becomes exothermic, and at ~800 K, glowing combustion begins. At this point also, a complex mixture of tar and gas products are released, which, when diluted with air, form a flammable mixture. When this mixture ignites, flaming combustion occurs, which converts the complex mixture of relatively reduced substances emitted during pyrolysis to simple molecules, particularly CO_2 , H_2O , NO , N_2O , N_2 , and SO_2 . Depending on the interaction between chemical kinetics and physical dynamics in the flame, intermediate products of flame radical chemistry, like CO , CH_4 , H_2 , C_2H_4 , C_2H_2 , PAH, and soot particles, are also released during this stage.

When most volatiles have been released from the near-surface region of the fuel, flaming combustion ceases, and smoldering begins. The energy-producing reaction in this phase is the gas-solid reaction between oxygen and carbon in the char layer at the fuel surface, yielding predominantly CO at temperatures usually <850 K. This lower-temperature process emits large amounts of incompletely oxidized pyrolysis products that are similar to the products of the initial solid phase decomposition [Lobert *et al.*, 1991; Yokelson *et al.*, 1997]. It is thus this mode of combustion that is responsible for the vast diversity of emission products that is shown in Table 1. The amount of substances emitted from a given fire and their relative proportions are thus determined to a large extent by the ratio of flaming to smoldering combustion. It is this proportion, often expressed as “combustion efficiency” (i.e., the ratio between CO_2 and the sum of all carbon species emitted), which has been used as an effective predictor for the emission of smoke gas composition from biomass fires [Ward *et al.*, 1996].

Open vegetation fires are typically dynamic fires, in which a moving fire front passes through a fuel bed, such as a savanna or forest. Consequently, all combustion types are present at any given time, and their combined emissions are released into the smoke plume. Their proportions vary over time, typically dominated by flaming in the earlier part of the fire and smoldering during the later part. Especially in forest fires, this late smoldering stage can continue for days or even weeks. This temporal separation of combustion types may introduce a bias toward flaming combustion products into emission measurements, as these late emissions may not always be sampled, especially in airborne campaigns. Time course studies on boreal and tropical forest fires show, however, that the vast majority of emissions (>99%) take place within ~4–8 hours after ignition, so that this bias is likely to be minor [Ward and Hardy, 1991; Ward *et al.*, 1992].

It may seem desirable to provide separate emission factors for flaming and smoldering combustion in Table 1, but this is not possible given the data available. Airborne measurements sample an integrated mixture of the emissions from both combustion types, and even at the ground the separation is incomplete because both processes occur simultaneously in a given patch at most times. This is not a serious problem, however, since each of the types of vegetation fires differentiated in Table 1 has a characteristic fuel composition and structure and typical ways in which these fires occur naturally or are managed by people; they tend to have characteristic ratios of flaming and smoldering combustion, which to a large extent determine their pattern of emission factors. In savanna fires, for example, flaming combustion dominates, and the emission factors for reduced species are fairly low. Charcoal making, however, is almost exclusively a smoldering and glowing process. Even in a given fire type, however, the proportion of flaming combustion can vary considerably as a function of internal parameters, particularly fuel moisture and structure, and external parameters, such as the movement of the flame front relative to wind direction or terrain slope. As a result, the emission pattern

Table 1. Emission Factors for Pyrogenic Species Emitted From Various Types of Biomass Burning^a

Species	Savanna and Grassland ^b	Tropical Forest ^c	Extratropical Forest ^d	Biofuel Burning ^e	Charcoal Making ^f	Charcoal Burning ^g	Agricultural Residues ⁱ
CO ₂	1613 ± 95	1580 ± 90	1569 ± 131	1550 ± 95	440	2611 ± 241	1515 ± 177
CO	65 ± 20	104 ± 20	107 ± 37	78 ± 31	70	200 ± 38	92 ± 84
CH ₄	2.3 ± 0.9	6.8 ± 2.0	4.7 ± 1.9	6.1 ± 2.2	10.7	6.2 ± 3.3	2.7
Total nonmethane hydrocarbons	3.4 ± 1.0	8.1 ± 3.0	5.7 ± 4.6	7.3 ± 4.7	2.0	2.7 ± 1.9	(7.0) ^h
C ₂ H ₂	0.29 ± 0.27	0.21–0.59	0.27 ± 0.09	0.51–0.90	0.04	0.05–0.13	(0.36) ^h
C ₂ H ₄	0.79 ± 0.56	1.0–2.9	1.12 ± 0.55	1.8 ± 0.6	0.10	0.46 ± 0.33	(1.4) ^h
C ₂ H ₆	0.32 ± 0.16	0.5–1.9	0.60 ± 0.15	1.2 ± 0.6	0.10	0.53 ± 0.48	(0.97) ^h
C ₃ H ₄	0.022 ± 0.014	0.013	0.04–0.06	(0.024) ^h	—	(0.06) ^h	(0.032) ^h
C ₃ H ₆	0.26 ± 0.14	0.55	0.59 ± 0.16	0.5–1.9	0.06	0.13–0.56	(1.0) ^h
C ₃ H ₈	0.09 ± 0.03	0.15	0.25 ± 0.11	0.2–0.8	0.04	0.07–0.30	(0.52) ^h
1-butene	0.09 ± 0.06	0.13	0.09–0.16	0.1–0.5	—	0.02–0.20	(0.13) ^h
i-butene	0.030 ± 0.012	0.11	0.05–0.11	0.1–0.5	—	0.01–0.16	(0.08) ^h
trans-2-butene	0.024 ± 0.014	0.05	0.01–0.05	0.05–0.3	—	0.01–0.06	(0.04) ^h
cis-2-butene	0.021 ± 0.011	0.042	0.008–0.13	0.05–0.18	—	0.01–0.03	(0.05) ^h
Butadiene	0.07 ± 0.05		0.06–0.08	0.11–0.36	—	0.01–0.10	(0.09) ^h
n-butane	0.019 ± 0.09	0.041	0.069 ± 0.038	0.03–0.13	—	0.02–0.10	(0.06) ^h
i-butane	0.006 ± 0.003	0.015	0.022 ± 0.009	0.01–0.05	—	0.006–0.01	(0.015) ^h
1-pentene	0.022 ± 0.010	0.056	0.04–0.07	0.5	—	0.028	0.008
n-pentane	0.005 ± 0.004	0.014	0.05–0.06	0.07	—	0.10	(0.025) ^h
2-methyl-butenes	0.008 ± 0.004	0.074	0.033	0.16	—	0.015	0.007
2-methyl-butane	0.011 ± 0.012	0.008	0.026–0.029	0.08	—	0.07	(0.018) ^h
Isoprene	0.020 ± 0.012	0.016	0.10	0.15–0.42	—	0.017	(0.05) ^h
Cyclopentene	0.012 ± 0.008	(0.02) ^h	0.019	0.61	—	0.035	(0.02) ^h
4-methyl-1-pentene	0.048	0.048	(0.05) ^h	0.015	—	(0.09) ^h	0.016
1-hexene	0.037 ± 0.016	0.063	0.07–0.11	(0.05) ^h	—	(0.13) ^h	0.013
n-hexane	0.039 ± 0.045	(0.05) ^h	0.03–0.06	(0.04) ^h	—	0.063	(0.05) ^h
Isohexanes	0.05	(0.08) ^h	(0.08) ^h	(0.06) ^h	—	(0.15) ^h	(0.08) ^h
Heptane	0.05	(0.08) ^h	(0.08) ^h	(0.06) ^h	—	(0.15) ^h	(0.08) ^h
Octenes	0.003–0.008	0.012	0.005	(0.007) ^h	—	(0.017) ^h	0.004
Terpenes	0.015	(0.15) ⁱ	0.22	(0.15) ⁱ	—	0.0	(0.015) ^h
Benzene	0.23 ± 0.11	0.39–0.41	0.49 ± 0.08	1.9 ± 1.0	—	0.3–1.7	0.14
Toluene	0.13 ± 0.06	0.21–0.29	0.40 ± 0.10	1.1 ± 0.7	—	0.08–0.61	0.026
Xylenes	0.045 ± 0.025	0.04–0.08	0.20	0.55 ± 0.44	—	0.04–0.22	0.01
Ethylbenzene	0.013 ± 0.003	0.013–0.035	0.048	0.17 ± 0.12	—	0.01–0.07	0.03
Styrene	0.024	(0.03) ^h	0.13	0.04–0.5	—	0.03–0.22	(0.03) ^h
PAH	0.0024	(0.025) ^j	(0.025) ^j	(0.025) ^j	—	(0.025) ^j	(0.025) ^j
Methanol	(1.3) ^h	(2.0) ^h	2.0 ± 1.4	(1.5) ^h	0.16	(3.8) ^h	(2.0) ^h
Ethanol	(0.011) ^h	(0.018) ^h	0.018	(0.013) ^h	—	(0.03) ^h	(0.018) ^h
1-propanol	0.025	(0.04) ^h	(0.04) ^h	(0.03) ^h	—	(0.08) ^h	(0.04) ^h
Butanols	0.008	0.009	(0.011) ^h	(0.008) ^h	—	(0.02) ^h	0.012
Cyclopentanol	0.032	0.031	(0.04) ^h	(0.03) ^h	—	(0.08) ^h	0.017
Phenol	0.003	0.006	(0.005) ^h	(0.004) ^h	—	(0.01) ^h	0.001
Formaldehyde	0.26–0.44	(1.4) ^h	2.2 ± 0.5	0.13 ± 0.05	—	(2.6) ^h	(1.4) ^h
Acetaldehyde	0.50 ± 0.39	(0.65) ^h	0.48–0.52	0.14 ± 0.05	—	(1.2) ^h	(0.65) ^h
Acrolein, propenal	0.08	(0.18) ^h	0.13–0.35	0.01–0.1	—	(0.35) ^h	(0.18) ^h
Propanal	0.009	(0.08) ^h	0.03–0.25	0.02–0.03	—	(0.15) ^h	(0.08) ^h
Butanals	0.053	0.071	0.21	0.04–0.05	—	(0.20) ^h	0.021
Hexanals	0.002–0.024	0.031	0.02	0.004–0.009	—	(0.04) ^h	0.012
Heptanals	0.003	0.003	(0.004) ^h	(0.003) ^h	—	(0.008) ^h	0.001
Acetone	0.25–0.62	(0.62) ^h	0.52–0.59	0.01–0.04	0.02	(1.2) ^h	(0.63) ^h
2-butanone	0.26	(0.43) ^h	0.17–0.74	0.03–0.06	—	(0.83) ^h	(0.44) ^h
2,3-butanedione	(0.57) ^h	(0.92) ^h	0.35–1.5	(0.68) ^h	—	(1.8) ^h	(0.9) ^h
Pantanones	0.01–0.02	0.028	0.09	(0.04) ^h	—	(0.09) ^h	0.007
Heptanones	0.006	0.002	(0.005) ^h	(0.004) ^h	—	(0.01) ^h	0.002
Octanones	0.015	0.019	(0.02) ^h	(0.016) ^h	—	(0.04) ^h	(0.02) ^h
Benzaldehyde	0.029	0.027	(0.036) ^h	0.02–0.03	—	(0.07) ^h	0.009
Furan	0.095	(0.48) ^h	0.40–0.45	0.65	—	(0.9) ^h	(0.5) ^h
2-methyl-furan	0.044–0.048	0.17	0.47	(0.18) ^h	—	(0.46) ^h	0.012
3-methyl-furan	0.006–0.011	0.029	0.05	(0.023) ^h	—	(0.06) ^h	0.003
2-ethylfuran	0.001	0.003	0.006	(0.003) ^h	—	(0.007) ^h	0.001
2,4-dimethyl-furan	0.008	0.024	(0.019) ^h	(0.014) ^h	—	(0.04) ^h	0.002
2,5-dimethyl-furan	0.002	(0.03) ^h	0.05	(0.021) ^h	—	(0.05) ^h	(0.03) ^h
Tetrahydrofuran	0.016	0.016	(0.02) ^h	(0.015) ^h	—	(0.04) ^h	0.006
2,3-dihydrofuran	0.012	0.013	(0.017) ^h	(0.012) ^h	—	(0.031) ^h	0.005
Benzofuran	0.014	0.015	0.026	(0.016) ^h	—	(0.04) ^h	0.004
Furfural	(0.23) ^h	(0.37) ^h	0.29–0.63	0.22	0.12	(0.72) ^h	(0.37) ^h
Methyl formate	(0.015) ^h	(0.025) ^h	0.025	(0.018) ^h	—	(0.05) ^h	(0.025) ^h

Table 1. (continued)

Species	Savanna and Grassland ^b	Tropical Forest ^c	Extratropical Forest ^d	Biofuel Burning ^e	Charcoal Making ^f	Charcoal Burning ^g	Agricultural Residues ⁱ
Methyl acetate	0.055	(0.10) ^h	0.09–0.12	(0.07) ^h	—	(0.19) ^h	(0.10) ^h
Acetonitrile	0.11	(0.18) ⁱ	0.19	(0.18) ⁱ	—	(0.18) ^h	(0.18) ^h
Formic acid	(0.7) ^h	(1.1) ^h	2.9 ± 2.4	0.13	0.20	(2.0) ^h	0.22
Acetic acid	(1.3) ^h	(2.1) ^h	3.8 ± 1.8	0.4–1.4	0.98	(4.1) ^h	0.8
H ₂	0.97 ± 0.38	3.6–4.0	1.8 ± 0.5	(1.8) ^h	—	(4.6) ^h	(2.4) ^h
NO _x (as NO)	3.9 ± 2.4	1.6 ± 0.7	3.0 ± 1.4	1.1 ± 0.6	0.04	3.9	2.5 ± 1.0
N ₂ O	0.21 ± 0.10	(0.20) ⁱ	0.26 ± 0.07	0.06	0.03	(0.20) ⁱ	0.07
NH ₃	0.6–1.5	(1.30) ⁱ	1.4 ± 0.8	(1.30) ⁱ	0.09	(1.30) ⁱ	(1.30) ⁱ
HCN	0.025–0.031	(0.15) ⁱ	(0.15) ⁱ	(0.15) ⁱ	(0.15) ⁱ	(0.15) ⁱ	(0.15) ⁱ
N ₂	(3.1) ^j	(3.1) ^j	(3.1) ^j	(3.1) ^j	—	(3.1) ^j	(3.1) ^j
SO ₂	0.35 ± 0.16	0.57 ± 0.23	1.0	0.27 ± 0.30	—	(0.40) ⁱ	(0.40) ⁱ
COS	0.015 ± 0.009	(0.04) ⁱ	0.030–0.036	(0.04) ⁱ	(0.04) ⁱ	(0.04) ⁱ	0.065 ± 0.077
CH ₃ Cl	0.075 ± 0.029	0.02–0.18	0.050 ± 0.032	0.04–0.07	(0.01) ⁱ	0.012	0.24 ± 0.14
CH ₃ Br	0.0021 ± 0.0010	0.0078 ± 0.0035	0.0032 ± 0.0012	(0.003) ⁱ	(0.003) ⁱ	(0.003) ⁱ	(0.003) ⁱ
CH ₃ I	0.0005 ± 0.0002	0.0068	0.0006	(0.001) ⁱ	—	(0.001) ⁱ	(0.001) ⁱ
Hg ⁰	0.0001	(0.0001) ⁱ	(0.0001) ⁱ	(0.0001) ⁱ	—	(0.0001) ⁱ	(0.0001) ⁱ
PM _{2.5}	5.4 ± 1.5	9.1 ± 1.5	13.0 ± 7.0	7.2 ± 2.3	—	(9) ⁱ	3.9
TPM	8.3 ± 3.2	6.5–10.5	17.6 ± 6.4	9.4 ± 6.0	4.0	(12) ⁱ	13
TC	3.7 ± 1.3	6.6 ± 1.5	6.1–10.4	5.2 ± 1.1	—	6.3	4.0
OC	3.4 ± 1.4	5.2 ± 1.5	8.6–9.7	4.0 ± 1.2	—	4.8	3.3
BC	0.48 ± 0.18	0.66 ± 0.31	0.56 ± 0.19	0.59 ± 0.37	—	1.5	0.69 ± 0.13
Levoglucosan	(0.28) ⁱ	0.42	(0.75) ⁱ	(0.32) ⁱ	—	—	(0.27) ⁱ
K	0.34 ± 0.15	0.29 ± 0.22	0.08–0.41	0.05 ± 0.01	—	0.40	0.13–0.43
CN	(3.4 × 10 ¹⁵) ^j	—	(3.4 × 10 ¹⁵) ^j	(3.4 × 10 ¹⁵) ^j			
CCN [at 1% SS]	(2 × 10 ¹⁵) ⁱ	(2 × 10 ¹⁵) ⁱ	[2.6 ± 4.2] × 10 ¹⁵	(2 × 10 ¹⁵) ⁱ	—	(2 × 10 ¹⁵) ⁱ	(2 × 10 ¹⁵) ⁱ
N _(>0.12 μm diam)	1.2 × 10 ¹⁵	(1 × 10 ¹⁵) ⁱ	(1 × 10 ¹⁵) ⁱ	(1 × 10 ¹⁵) ⁱ	—	(1 × 10 ¹⁵) ⁱ	(1 × 10 ¹⁵) ⁱ

^aEmission factors are given in gram species per kilogram dry matter burned. See text for the conventions used for reporting uncertainties. Abbreviations are as follows: PM_{2.5}, particulate matter <2.5 μm diameter; TPM, total particulate matter; TC, total carbon; BC, black carbon; CN, condensation nuclei; CCN, cloud condensation nuclei at 1% supersaturation; and N_(>0.12 μm diam), particles > 0.12 μm diameter. Values in parentheses represent estimates for emission factors that have not been measured directly. Estimation methods are indicated by superscripts.

^bData sources are Delmas and Servant [1982], Brunke et al. [2001], Cofer et al. [1988, 1989], Ward and Hardy [1989], Cofer et al. [1990a], Bonsang et al. [1991], Delmas et al. [1991], Ward et al. [1991], Laursen et al. [1992], Ward et al. [1992], Hao and Ward [1993], Lacaux et al. [1993], Hurst et al. [1994a, 1994b], Manö and Andreae [1994], Singh et al. [1994], Bonsang et al. [1995], Cachier et al. [1995], Echalar et al. [1995], Gaudichet et al. [1995], Helas et al. [1995], Lacaux et al. [1995], Masclet et al. [1995], Nguyen et al. [1995], Rudolph et al. [1995], Scholes [1995], Anderson et al. [1996], Andreae et al. [1996a, 1996b], Blake et al. [1996], Cachier et al. [1996], Cofer et al. [1996a], Hao et al. [1996a, 1996b], Koppmann et al. [1996], Lacaux et al. [1996], Le Canut et al. [1996], Ward et al. [1996], Lee et al. [1997], Andreae et al. [1998], Ferek et al. [1998], Friedli et al. [2001], and Yamasoe et al. [2000].

^cData sources are Greenberg et al. [1984], Andreae et al. [1988], Ward et al. [1991, 1992], Hao and Ward [1993], Delmas et al. [1995], Andreae et al. [1996b], Blake et al. [1996], Koppmann et al. [1996], Andreae et al. [1998], Ferek et al. [1998], Yamasoe et al. [2000], and Graham et al. [2001].

^dData sources are Miner [1969], Eagan et al. [1974], Crutzen et al. [1979], Stith et al. [1981], Ward and Hardy [1986], Radke et al. [1988], Cofer et al. [1989, 1990b], Hegg et al. [1990], Radke et al. [1990], Susott et al. [1990], Ward et al. [1990], Cofer et al. [1991], Einfeld et al. [1991], Radke et al. [1991], Ward and Hardy [1991], Laursen et al. [1992], Wofsy et al. [1992], Hao and Ward [1993], Nance et al. [1993], Blake et al. [1994], Lefer et al. [1994], Manö and Andreae [1994], Singh et al. [1994], Manö [1995], Cofer et al. [1996a, 1996b], FIRESCAN Science Team [1996], Hobbs et al. [1996], Hurst et al. [1996], Martins et al. [1996], Vose et al. [1996], Worden et al. [1997], Andreae et al. [1998], Cofer et al. [1998], Yokelson et al. [1999], Friedli et al. [2001], and Goode et al. [2000].

^eData sources are Rasmussen et al. [1980], Smith et al. [1993], Brocard et al. [1996], Liousse et al. [1996], Zhang and Smith [1996], Zhang et al. [1999], Kituyi et al. [2001], and Veldt (unpublished manuscript, 1992).

^fData sources are Crutzen et al. [1979], De Angelis et al. [1980], Myers [1980], Rasmussen et al. [1980], Dasch [1982], Edgerton et al. [1986], Smith [1988], Delmas et al. [1991], Joshi [1991], Veldt [1992], Hao and Ward [1993], Smith et al. [1993], Brocard et al. [1996], Cachier et al. [1996], Liousse et al. [1996], Piccot et al. [1996], Zhang and Smith [1996], Schauer [1998], Zhang and Smith [1999], Zhang et al. [1999], Kituyi et al. [2001], Ludwig et al. [2001], Zhang et al. [2000], and Veldt (unpublished manuscript, 1992).

^gData sources are Rasmussen et al. [1980], Lobert et al. [1991], Hao and Ward [1993], Nguyen [1994], Nguyen et al. [1995], Andreae et al. [1996a], Koppmann et al. [1996], Liousse et al. [1996], Zhuang et al. [1996], Andreae et al. [1998], de Zarate et al. [2000], Kituyi et al. [2001], and Ludwig et al. [2001].

^hExtrapolation is based on emission ratios to CO.

ⁱValue is best guess.

^jEstimate is based on laboratory studies.

from a particular fire can be quite different from the average values presented in Table 1.

3.2. Emission Factors for Chemical Species From Fires in Various Vegetation Types or Burning Practices

During the numerous fire emission studies over the last few decades, including the various Biomass Burning Experiment (BIBEX) field experiments, a large number of emission ratios or factors have been obtained. However, there is no coherent, critical compilation of these data on which a table of "recommended" emission values could be based. Previous reviews have either only provided average values for "biomass burning" in general, ignoring differences between the different types of fires [e.g., *Andreae*, 1993], or given values only for specific fire types or regions, such as savanna fires [e.g., *Andreae*, 1997]. In Table 1 we present emission data for the most important types of fire regimes (savannas and grasslands, tropical forest, extratropical forest, domestic biofuel burning, charcoal production and combustion, and agricultural waste burning) for a wide variety of gaseous and particulate emission products. These emission factors are based on an analysis of some 130 publications.

One problem that we encountered in compiling Table 1 was that the amount of information available for each data cell was quite different between chemical species and fire types. For the major carbon species emitted from fires, such as CO₂, CO, and CH₄, sufficient data are available for all fire types. Conversely, adequate data exist for most species listed in Table 1 in the case of savanna and grassland fires, but for other fire types only the emissions of some key species have been determined satisfactorily. In order to represent the different levels of information upon which the estimates in Table 1 are founded, we have adopted the following convention: When three or more values (based on independent references) are available for a given cell, the results are given as means and standard deviations ($x \pm s$). In the case of two available measurements they are given as a range, and where only a single measurement is available, it is given without an uncertainty estimate. For single measurements it can usually be assumed that the uncertainty is no less than a factor of 3.

For combinations of species and fire type without data available, we have supplied estimates (in parentheses in Table 1) using the following three techniques of extrapolation. (1) For species not closely tied to the smoldering stage and where information is available for most but not all fire types, a weighted estimate was calculated from the averages given for those fire types where data are available. The weighting was based on the global amount of carbon burned in each fire type (see Table 2). (2) The emission factors of species that are predominantly emitted in the smoldering phase can be estimated using the fact that they tend to be closely correlated to the emission factor for CO. For this purpose we calculated a mean ratio EF_X/EF_{CO} from the available data and multiplied it by the EF_{CO} characteristic for the fire types for which EF_X was not measured. (3) For the remaining cases, where there was inadequate information to derive an objective estimate, we provide a "best guess," based on the subjective evaluation of the available information, including the results from laboratory burning studies. For different ways of estimation see footnotes in Table 1.

As mentioned in section 3.1, the level of information available on the different fire types and chemical species varies dramatically across Table 1. While savanna fire emissions are reasonably well characterized, glaring deficiencies prevail for other fire types. An example is charcoal making: Here all the information available is based on the study of *Lacaux et al.* [1994], which did not include a large set of organic species. Similarly, the organic emissions from charcoal burning are characterized in only one study [*Zhang and*

Smith, 1996]. The remaining information on volatile organic compound emissions from charcoal making and burning is based on extrapolations provided by C. Veldt (GEIA note on residential biomass burning: A short communication about emission factors for the residential combustion of biomass fuels, unpublished manuscript, 1992) (hereinafter referred to as Veldt, unpublished manuscript, 1992).

While the paucity of data on charcoal-related emissions may not be a major problem, given the relatively small contribution of charcoal burning to the global budget, the limited data on emissions from wood and dung burning have proven to cause serious problems in the interpretation of the measurements of aerosol and trace gas composition downwind of regions where domestic biomass burning is a major contributor, for example, India. Here the relative contributions of biomass and fossil fuel burning have proven difficult to assess, and emission inventories and ambient measurements have proven difficult to reconcile, not in the least because of the poor state of knowledge concerning emission factors [*Novakov et al.*, 2000; *Mayol-Bracero et al.*, 2001]. In India, particularly, dung cakes contribute significantly to the biofuel budget [*Streets and Waldhoff*, 1999]. Because of their high N, S, and Cl content they have high emission factors for species such as NO, NH₃, SO₂, and CH₃Cl. They have, however, not been included in the average given for biofuels in Table 1 because of the limited availability of emission data and because of the difficulty of assessing a broadly representative dung cake contribution to the fuel mix. Supplementary data on these fire types can be obtained from the authors.

We cannot discuss here all the deficiencies in the presently available data on pyrogenic emissions, and we invite the reader to examine Table 1 in detail. However, we would like to point out some key compounds of global interest for which more data are urgently needed. In this context the small number of field data on emissions of acetonitrile and other nitrogen species, as well as the organic acids and other oxygenized organic compounds, are particularly problematic. With the development of novel in situ techniques that allow the determination of these compounds without the problems associated with sample collection and storage, we can hope for improved data sets for these species in the near future [*Holzinger et al.*, 1999; *Crutzen et al.*, 2000; *Reiner et al.*, 2001]. There are several compounds and compound classes that have not been included in Table 1, particularly the amines and pyrroles. These compounds have been identified in laboratory studies of biomass burning but have not yet been measured in the field. Because of the difficulty of extrapolating from laboratory to field conditions, we have decided to exclude these compounds from the present analysis.

Because of the fairly low combustion temperatures in biomass fires (compared to fossil fuel combustion), atmospheric N₂ is not converted to fixed nitrogen to a significant extent, and the nitrogen species emissions are based only on the fuel nitrogen. For this reason, linear relationships have been found between N₂O [*Lobert et al.*, 1991] and NO_x [*Lacaux et al.*, 1993] emission and fuel nitrogen content. NO, NO₂, N₂O, and molecular N₂ are released predominantly during flaming combustion, whereas NH₃, amines, and nitriles are associated with smoldering combustion. The most abundant N species in the emissions is typically NO, and it has been frequently assumed that all NO_x is released in this form. Where NO₂ has been determined selectively, however, it has been found that it represents some 10% of NO_x already directly above the fire [*Griffith et al.*, 1991; *Yokelson et al.*, 1996]. Our emission factor estimate for N₂O is based only on data that have been collected without any or with minimal sample storage to avoid the positive artifact that results from prolonged storage in containers. NH₃ emissions dominate in the smoldering stage and can account for most of the N emissions in that phase [*Griffith et al.*, 1991]. This trace gas used to be considered a minor N species in smoke,

Table 2. Global Emission of Selected Pyrogenic Species in the Late 1990s, Based on the Emission Factors in Table 1 and the Biomass Burning Estimates of J. A. Logan and R. Yevich^a

	Savanna and Grassland	Tropical Forest	Extratropical Forests	Biofuel Burning	Charcoal Making	Charcoal Burning	Agricultural Residues ^b	Total
Dry matter burned, Tg	3,160	1,330	640	2,701	158	38	540	8,600
CO ₂	5,096	2,101	1,004	4,187	70	99	818	13,400
CO	206	139	68	209	11	7.6	50	690
CH ₄	7.4	9.0	3.0	16.5	1.7	0.24	1.5	39
NMHC	10.7	10.8	3.6	19.6	0.3	0.10	3.5	49
C ₂ H ₂	0.92	0.53	0.17	1.90	0.01	0.004	0.20	3.7
Methanol	3.8	2.6	1.3	3.9	0.02	0.14	0.9	12.7
Formaldehyde	1.1	1.8	1.4	0.4	—	0.10	0.7	5.5
Acetaldehyde	1.6	0.86	0.32	0.37	—	0.05	0.31	3.5
Acetone	1.4	0.83	0.35	0.06	0.00	0.05	0.30	3.0
Acetonitrile	0.33	0.24	0.12	0.49	—	0.007	0.10	1.3
Formic acid	2.1	1.4	1.8	0.36	0.03	0.08	0.1	5.9
Acetic acid	4.2	2.8	2.5	2.4	0.15	0.15	0.4	12.6
H ₂	3.1	5.1	1.2	4.8	—	0.17	1.1	15.3
NO _x	12.2	2.2	1.9	2.9	0.01	0.15	1.3	20.7
N ₂ O	0.67	0.27	0.17	0.16	0.00	0.008	0.04	1.31
NH ₃	3.4	1.7	0.88	3.5	0.01	0.05	0.7	10.3
HCN	0.09	0.20	0.10	0.41	0.02	0.006	0.08	0.90
N ₂	9.8	4.1	2.0	8.4	—	0.12	1.7	26.0
SO ₂	1.1	0.76	0.64	0.74	—	0.015	0.22	3.5
COS	0.05	0.05	0.02	0.11	0.01	0.002	0.03	0.27
CH ₃ Cl	0.24	0.10	0.03	0.14	—	0.0005	0.13	0.65
CH ₃ Br	0.006	0.010	0.002	0.008	—	0.00011	0.002	0.029
CH ₃ I	0.0016	0.0090	0.0004	0.0027	—	0.00004	0.0005	0.014
Hg ⁰	0.0003	0.0001	0.00006	0.0003	—	—	0.0001	0.0008
PM _{2.5}	16.1	12.0	8.3	19.4	—	0.34	2.1	58.3
TPM	26.2	11.3	11.3	25.5	0.63	0.46	7.0	82.4
TC	11.7	8.7	5.3	14.0	—	0.24	2.2	42.2
OC	10.6	7.0	5.8	10.7	—	0.18	1.8	36.1
BC	1.5	0.88	0.36	1.6	—	0.06	0.37	4.8
K	1.09	0.39	0.16	0.14	—	0.02	0.15	1.9
CN	1.1E + 28 ^c	4.5E + 27	2.2E + 27	9.2E + 27	—	1.3E + 26	1.8E + 27	2.9E + 28
CCN (1% SS)	6.3E + 27	2.7E + 27	1.7E + 27	5.4E + 27	—	7.6E + 25	1.1E + 27	1.7E + 28
N(_{<0.12} (μm diam))	3.7E + 27	1.3E + 27	6.4E + 26	2.7E + 27	—	3.8E + 25	5.4E + 26	9.0E + 27

^aValues are given in mass of species per year (Tg yr⁻¹). Biomass burning estimates are from R. Yevich (personal communication, 2001).

^bValue excludes agricultural waste used as biofuel.

^cRead, for example, 1.1E + 28 as 1.1×10^{28} .

but studies with improved analytical techniques, particularly Fourier transform infrared spectroscopy, have shown that it is often emitted at roughly equimolar amounts to NO_x [Griffith *et al.*, 1991; Yokelson *et al.*, 1996; Worden *et al.*, 1997; Yokelson *et al.*, 1997; Goode *et al.*, 1999]. Some 30–40% of the fuel nitrogen is released in the form of molecular N₂ [Kuhlbusch *et al.*, 1991]. Because it is impossible to detect the small enhancement of atmospheric N₂ that would be present in a smoke plume, only laboratory data are available for the emission factor of this species, and the value obtained in the single study published on N₂ emission has been used in Table 1. Hydrogen cyanide (HCN) has been proposed as a potentially valuable tracer of biomass burning since it can be determined by remote sensing from space [Li *et al.*, 2000]. Unfortunately, there are only two field studies during which HCN was measured, both in Australian savanna fires [Hurst *et al.*, 1994a, 1994b], and the emission factors from these studies are an order of magnitude lower than the results of laboratory emission measurements ($0.31 \pm 0.25 \text{ mg kg}^{-1}$) [Lobert *et al.*, 1991; Holzinger *et al.*, 1999]. In view of the large uncertainty regarding this interesting species, additional field measurements would be highly desirable.

The methyl halides, CH₃Cl, CH₃Br, and CH₃I, are formed predominantly in the smoldering stage, probably because of reaction between methanol and HCl etc. catalyzed at glowing char surfaces [Reinhardt and Ward, 1995] or by radical reactions in flames. Since the emission factors depend both on the halogen

content of the fuel and on the proportion of flaming to smoldering combustion [Andreae *et al.*, 1996b], considerable variability in the emission ratios has been observed.

Fairly good information is now available on aerosol emissions from most types of burning, with the exception of agricultural fires and charcoal use. This does not apply, however, to measurements of the number of particles emitted per amount of biomass burned. The determination of this parameter is problematic, since it changes very rapidly in the initial phase of plume development [Liouesse *et al.*, 1995; Hobbs *et al.*, 1996; Reid *et al.*, 1998]. Just outside of the flames, particle concentrations are very high (ten thousands to millions per cubic centimeter), and rapid reduction of particle numbers by coagulation takes place. This is a highly nonlinear process and slows down rapidly when the particle concentration drops because of coagulation and dilution. At the same time the particle size grows, both because of coagulation and condensation of vapors as the plume cools down. The ability of the particles to act as cloud condensation nuclei (CCN) also increases sharply during plume aging, presumably because of condensational uptake of water-soluble material on the particles [Hobbs *et al.*, 1996]. For these reasons the values in Table 1 given for particle number emission factors must be considered as fairly rough estimates. They are meant to apply to a slightly aged (1–2 hours) smoke plume to avoid the large temporal variations shortly after emission. The results for the different particle classes appear reasonably internally consistent, even though they come from a

variety of sources. The ratio of ~ 3 between total particle number and the size fraction $>0.12 \mu\text{m}$ is reasonably consistent with the number median diameters of $\sim 0.1 \mu\text{m}$ typically observed for slightly aged smoke [e.g., Hobbs *et al.*, 1996; Reid *et al.*, 1998]. Also, the CN/CCN (1%) ratio of ~ 1.7 agrees with the observation that biomass smoke particles tend to be good CCN after short aging [Warner and Twomey, 1967; Rogers *et al.*, 1991; Hobbs *et al.*, 1996; Pham-Van-Dinh *et al.*, 1996].

3.3. Emissions From Global Biomass Burning

While, as we have shown in section 3.2, the average emission factors for many important species, such as CO and CH_4 , are now known with an uncertainty of about 20–30%, large uncertainties persist for regional and global fire emissions because of the difficulties inherent in estimating the amount of biomass burned. The estimation of the amounts of biomass combusted per unit area and time is still based on rather crude assessments and has not yet benefited significantly from the remote sensing tools becoming available at this time. Where comparisons between different approaches (e.g., inventories versus remote sensing) and combustion estimates have been made, they have shown large differences for specific regions. One example comes from the savanna regions of southern Africa, where a difference of almost an order of magnitude has been found between regional estimates based on estimates of fire frequencies in the various vegetation types and those based on fire counts obtained from remote sensing [Scholes *et al.*, 1996; Scholes and Andreae, 2000].

Table 2 provides a set of global emission estimates for the late 1990s, based on the emissions factors in Table 1 and the biomass burning estimates of J. A. Logan and R. Yevich (R. Yevich, personal communication, 2001). The data on charcoal use were updated using the Food and Agriculture Organization's FAOSTAT database at <http://apps.fao.org/>). Uncertainties are not explicitly stated in Table 2, in part because there is not enough information to estimate them quantitatively. For each entry in Table 2 the appropriate error would result from error propagation from the emission factor data in Table 1 and the estimates of biomass burned. The uncertainties related to the emission factors have been discussed in section 3.2, where it was pointed out that in many cases there are not enough data available for a statistical error analysis. Given the methodological problems associated with estimating the amount of biomass burned, there is again no basis for a statistically valid error estimate. The inventory-based estimates for biomass burned have changed little over the last decade [Scholes *et al.*, 2001], but this is due more to the use of a relatively constant underlying information base and methodology than to actual accuracy of the data. Until tools become available to perform independent validation of these estimates, we must assume that they are uncertain to at least $\pm 50\%$.

Some support for the accuracy of our estimates comes from the application of inverse models to the analysis of the budgets of CO and CH_4 . Our estimate of pyrogenic CH_4 (39 Tg yr^{-1}) agrees very well with the inverse modeling estimate of $40 \pm 12 \text{ Tg yr}^{-1}$ by Hein *et al.* [1997]. For a variety of model scenarios an inverse model considering both CO concentration and isotopic composition data predicted CO emissions of $483\text{--}633 \text{ Tg yr}^{-1}$, $140\text{--}245 \text{ Tg yr}^{-1}$, and $0\text{--}87 \text{ Tg yr}^{-1}$ for tropical forest burning, savanna burning, and burning at latitudes $>30^\circ\text{N}$, respectively [Bergamasci *et al.*, 2000]. The total pyrogenic CO emission was estimated to be between 663 and 807 Tg yr^{-1} . Comparison with Table 2 shows very good agreement between our a priori estimate and the inversion results for the total CO source but shows poorer agreement for the individual contributions, especially that of forest burning. This is probably because Table 2 includes emission from fire types not included (biofuel burning, agricultural burning, and charcoal making) in the categories that are represented in the

model. While the agreement between our inventory-based estimates and the results of inverse modeling do not provide very narrow constraints on the accuracy of our estimates, they suggest that the pyrogenic emissions of CO and CH_4 lie within $\pm 50\%$ of the values in Table 2. On the basis of a comparison between modeled and observed CO concentrations, Galanter *et al.* [2000] also conclude that global estimates of pyrogenic emissions for these species are considerably more accurate than a factor of 2. Since the emission factors for CO and CH_4 are fairly well known, this implies that the total amount of biomass burned annually is also within an uncertainty range of about $\pm 50\%$ around the value of $8600 \text{ Tg dry matter}$ used in Table 2.

4. Conclusions

Considerable progress has been made over the last decade with regard to the determination of emission factors from biomass burning. A critical evaluation of the available data shows that a vast number of chemical species have been identified in biomass burning smoke and that reliable emission information exists for most of the key species, at least for savanna fires. There remain, however, serious gaps for important species, including ones that could be valuable atmospheric tracers, such as HCN and acetonitrile. Some combustion types also need further study, for example, the various types of biofuel use, including charcoal making. The global emission estimates from biomass burning have been refined but require further validation. This applies particularly to the estimates of biomass burned as a function of space, time, and type of combustion. The agreement between the results from inverse models and the inventory-based estimates presented here is encouraging, but more rigorous constraints of emission estimates could come from regional experiments (similar to that described by Cautenet *et al.* [1999]) that are designed to test the agreement between emission inventories and transport and chemistry models.

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